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Article

# Assessing the Potential for Ion Selective Electrodes and Dual Wavelength UV Spectroscopy as a Rapid on-Farm Measurement of Soil Nitrate Concentration

Rory Shaw <sup>1,\*</sup>, A. Prysor Williams <sup>1</sup>, Anthony Miller <sup>2</sup> and Davey L. Jones <sup>1</sup>

<sup>1</sup> School of Environment, Natural Resources and Geography, Bangor University, Gwynedd, LL57 2UW, UK; E-Mails: prysor.williams@bangor.ac.uk (A.P.W.); d.jones@bangor.ac.uk (D.L.J.)

<sup>2</sup> John Innes Centre, Norwich Research Park, Norwich, NR4 7UH, UK; E-Mail: Tony.Miller@jic.ac.uk

\* Author to whom correspondence should be addressed; E-Mail: rory.shaw@bangor.ac.uk; Tel.: +44-1248-382579; Fax: +44-1248-354997.

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**Abstract:** Current fertiliser recommendations for nitrogen are limited in their accuracy and may be improved by the use of simple on-farm soil rapid tests. This paper investigates the potential for using nitrate ( $\text{NO}_3^-$ ) ion selective electrodes (ISEs) and dual wavelength UV spectroscopy as part of a rapid soil  $\text{NO}_3^-$  diagnostic test. Three soil types, representing the major soil types for agriculture in the western UK, were tested. For the three soils, the ISE rapid test procedure gave a near 1:1 response ( $r^2 = 0.978, 0.968, 0.989$ ) compared to the internationally-approved standard laboratory method. However, the accuracy of the ISE rapid test was reduced at low soil  $\text{NO}_3^-$  concentrations ( $<10 \text{ mg NO}_3^- \text{ L}^{-1}$ ). We also show that  $\text{NO}_3^-$  analysis of  $\text{H}_2\text{O}$  soil extracts by dual wavelength UV spectroscopy was also highly correlated ( $r^2 = 0.978, 0.983, 0.991$ ) to the standard laboratory method. We conclude that both ISE and dual wavelength UV spectroscopy have clear potential to be used for the rapid on-farm determination of soil  $\text{NO}_3^-$  concentration. Barriers to use of these field-based assessment tools include, farmer perception of cost-benefit, general attitude to new technologies and the ability to generate useful fertiliser use strategies from soil  $\text{NO}_3^-$  measurements.

**Keywords:** crop nutrients; fertiliser management; nitrogen use efficiency; soil analysis

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## 1. Introduction

Improving nitrogen use efficiency (NUE) is a major goal within agricultural systems [1] and is key to the success of sustainable intensification [2]. Use of nitrogen (N) fertilisers represents the major N input in most farming systems and both under- and over-use of N fertilisers can represent an economic loss for the farmer, while over-use may cause significant environmental pollution [3–6]. In a purely economic sense, an optimum N fertiliser strategy can be defined as the point at which the cost of an additional unit of N is no longer covered by the resulting increase in crop yield. Defining an environmental optimum rate of N addition, however, is much more problematic. Calculating an optimum N fertilisation strategy is extremely desirable, but very difficult to achieve due to the inherent complexity of the soil-plant system, temporal and spatial variability and the importance of uncontrolled variables such as weather [5]. One theoretical method for improving NUE is to ensure synchronicity of supply and demand, both spatially and temporally, by maintaining the pool of plant available N in the soil at the minimum size required to meet crop demand [7,8].  $\text{NO}_3^-$  is typically the most important crop-available form of N in most temperate climate, near neutral pH soils, although in some grassland soils ammonium ( $\text{NH}_4^+$ ) may dominate. The mobility of  $\text{NO}_3^-$  within soil makes it easy for plants to uptake, but this property also makes it prone to being leached from the soil profile when field capacity is reached, with resulting water pollution issues. Regular testing of soil  $\text{NO}_3^-$  concentration, over the course of the growing season, may help farmers improve their nutrient management strategy by better matching supply and demand.

Current methods of calculating the N requirement of the crop over the growing season, and hence fertiliser N additions, require a prediction of the crop yield, based on soil type, climatic zone, topography and other variables. The amount of crop-available N that can be supplied by the soil in its pre-fertilised state over the growing season is then measured or estimated; this is known as the soil nitrogen supply (SNS). The difference between crop requirement and the SNS can then be made up for by addition of N fertiliser [5]. A variation of this method is widely used in the UK and is prescribed by The Fertiliser Manual RB209 [9]. A key component of the SNS is the concentration of soil mineral nitrogen (SMN), which consists of  $\text{NH}_4^+$  and  $\text{NO}_3^-$ . RB209 provides tables to allow estimation of a field's SNS depending upon its soil type and previous management. This estimation can also be supplemented by measuring the SMN of the pre-fertilised soil.

Traditionally, laboratory analysis has been used for soil testing. However, it is expensive and time-consuming and therefore not suitable as a method of regular and frequent testing. There are also other problems with laboratory analysis. Unwanted mineralisation and nitrification/denitrification of the samples may occur during transport and storage prior to analysis. Significant changes to the intrinsic soil  $\text{NO}_3^-$  status may also occur, due to changes in the prevailing weather (e.g., during the delay between sampling and receiving the results). Farmer-operated rapid diagnostic testing potentially offer a cheap and instantaneous determination of soil  $\text{NO}_3^-$  status where the results can be used to directly inform nutrient management strategies, benefiting both the farmer and the environment. Previous work on rapid soil tests have largely been based on colorimetric strips combined with a handheld reflectometer [10–13] and ion selective electrodes (ISEs) [12,14], which have been described as semi-quantitative [15]. New, more quantitative methods are therefore required.

This study aimed to evaluate two contrasting rapid test methods for evaluating soil  $\text{NO}_3^-$  concentration. Firstly, we compared a rapid extraction method, which could be used in-field for the extraction of soil  $\text{NO}_3^-$ , coupled with  $\text{NO}_3^-$  determination using an ISE, to the standard laboratory determination of soil  $\text{NO}_3^-$ . Secondly, we evaluated the use of UV spectroscopy for  $\text{NO}_3^-$  determination in soil extracts in comparison to the ISE approach. The results were used to evaluate the potential of the two approaches for the on-farm measurement of soil  $\text{NO}_3^-$  status.

## 2. Materials and Methods

### 2.1. Soil Type and Sampling

Three contrasting soils were collected from Henfaes Research Station, Abergwyngregyn, UK ( $53^\circ 14' \text{N}$ ,  $4^\circ 01' \text{W}$ ). Soil 1 is a lowland, clay loam textured Eutric Cambisol collected from an area of no vegetation cover, which had been used for potato production the previous season. Soil 2 is a lowland, silty loam textured Dystric Gleysol collected from a poorly draining area of an intensively sheep grazed field (ca.  $>10$  ewe  $\text{ha}^{-1}$ ) receiving regular fertiliser inputs ( $120 \text{ kg N ha}^{-1} \text{ yr}^{-1}$ ) and dominated by *Lolium perenne* L. Soil 3 is a sandy loam textured Haplic Podzol collected from an upland, extensively grazed ( $<0.1$  ewe  $\text{ha}^{-1}$ ) unimproved acid grassland (*Pteridium aquilinum* L. Kuhn. and *Festuca ovina* L.). Prior to sampling, the overlying vegetation cover was removed and the soil sampled from a depth of 3–15 cm. After collection, the soil was stored in gas permeable polyethylene bags for immediate transport to the laboratory. The soil was refrigerated at  $<5^\circ \text{C}$  until it was needed for the experimental procedure. Immediately prior to use, the soil was sieved to 8 mm to remove large stones, roots, vegetation and earthworms and then thoroughly mixed.

### 2.2. Background Soil Analysis

Soil pH and electrical conductivity were determined in a 1:2 (w:v) soil:water mix using standard electrodes. Moisture content was determined by drying for 24 h at  $105^\circ \text{C}$ . Organic matter content was determined by loss-on-ignition at  $450^\circ \text{C}$  for 12 h. Total C and N were determined with a CHN2000 analyser (Leco Corp., St Joseph, MI, USA). Results for background soil analysis can be found in Table 1.

**Table 1.** Background characteristics of the three soils used in the experiments.

| Soil            | pH   | EC ( $\mu\text{S cm}^{-1}$ ) | Moisture content ( $\text{g g}^{-1}$ ) | Total C (%) | Total N (%) |
|-----------------|------|------------------------------|--|-------------|-------------|
| Eutric Cambisol | 6.53 | 59.4                         | 0.25                                   | 3.5         | 0.29        |
| Dystric Gleysol | 6.53 | 59.4                         | 0.28                                   | 1.1         | 0.10        |
| Haplic Podzol   | 5.34 | 12.9                         | 0.70                                   | 6.1         | 0.57        |

### 2.3. Ion Selective Electrode (ISE)

A commercially available  $\text{NO}_3^-$  ISE (ELIT 8021) with a solid state PVC polymer matrix membrane was used in conjunction with a double junction lithium acetate reference electrode (RE) (ELIT 003n), supplied by Electro Analytical Instruments (EAI) (Wembley, UK). The  $\text{NO}_3^-$  ISE is reported to have an operational concentration range from 0.3–6300  $\text{mg NO}_3^- \text{ L}^{-1}$ , a response time of  $<10$  s, working pH

range of pH 2–11, operational temperature range from 0–50 °C and an electrode slope of  $54 \pm 5$  mV decade<sup>-1</sup> at 25 °C. The ISE and RE were coupled with a multi-channel analyser (6+6-Channel Ion/pH/ORP/Tmp.Monitor MCC-SYSTi-6+6b) and corresponding PC software (MCC-MON-6+6c, Version 2.1.1) supplied by Electro Analytical Instruments (EAI) (Wembley, UK). Prior to initial use, the ISE was pre-conditioned in a 1000 mg L<sup>-1</sup> NO<sub>3</sub><sup>-</sup> solution for 4 h. The calibration is calculated and stored by the software using a semi-logarithmic interpolation method.

#### 2.4. NO<sub>3</sub><sup>-</sup> Determination Using Ion Selective Electrode Rapid Test Method

Before each set of measurements, the ISE was calibrated using a range of NO<sub>3</sub><sup>-</sup> solutions (1000, 100, 10, 1, 0.5 mg NO<sub>3</sub><sup>-</sup> L<sup>-1</sup>). The temperature of the calibrating solutions differed from the experimental measurements by a maximum of  $\pm 2$  °C. 10 g soil ( $n = 3$  for each soil type) was placed in a 50 cm<sup>3</sup> polypropylene tube and spiked with 1 mL of NO<sub>3</sub><sup>-</sup> solution (2000, 1800, 1600, 1400, 1200, 1000, 800, 600, 400, 200, 100 or 0 mg L<sup>-1</sup> (in addition the Eutric Cambisol was spiked with 20 and 10 mg L<sup>-1</sup>)) to achieve a range of intrinsic NO<sub>3</sub><sup>-</sup> concentrations reflective of those that might occur in the field. Extraction was then performed by the addition of 20 mL of double distilled (DD) H<sub>2</sub>O followed immediately by manual shaking by hand for 2 min. This extraction procedure is referred to as the rapid extraction method. The ISE was placed into the resulting soil slurry and a reading taken after 3 min. Between each measurement, the electrodes were rinsed with DD H<sub>2</sub>O and dried with paper tissue. The soil slurry was subsequently centrifuged (20 min at 4000 rev min<sup>-1</sup> followed by 20 min at 14,000 rev min<sup>-1</sup>) and the supernatant decanted for NO<sub>3</sub><sup>-</sup> analysis by the colorimetric Griess reaction method of Miranda *et al.* [16]—referred to as the standard laboratory method. This analysis reflects internationally accepted protocols for *ex situ* soil nitrate analysis [17].

It should be noted that ISEs respond not to the concentration but the activity of the specified ion. The activity of the ion depends upon both the concentration of the ion and the total ionic strength of the solution. There is little difference in concentration and activity at low ionic solution strength *i.e.* below 1 mM. However, above this they diverge leading to the potential for systematic error. In previous experiments using ISEs for determination of soil NO<sub>3</sub><sup>-</sup> concentration, ionic strength adjustment buffers (ISAB) have been used to keep the ionic strength of the calibrating solutions equal and approximately matched to samples being tested [18]. This was not possible in this experiment as adding NO<sub>3</sub><sup>-</sup> to the soil in varying amounts intrinsically meant that the ionic strength of the soil solution would be different between different treatments. The extracted solutions of soils amended with 1 mL of 2000 mg L<sup>-1</sup> NO<sub>3</sub><sup>-</sup>, would not exceed an ionic strength of 3 mM, which would result in a 6% difference between concentration and activity for monovalent ions. Whilst this has the potential to cause systematic error, this is partially offset by the fact that the ISEs were calibrated for concentration without ISAB. For the soils which were not spiked with NO<sub>3</sub><sup>-</sup>, the ionic strengths of the measured extractions did not exceed 1 mM (equivalent to an electrical conductivity of 120  $\mu$ S cm<sup>-1</sup>), which would not cause a significant error.

### 2.5. Nitrate Extraction and Determination by the Standard Lab Method

$\text{NO}_3^-$  was extracted from the soil using 1 M KCl or DD  $\text{H}_2\text{O}$  (10 g soil:20 mL) by mechanical shaking at  $150 \text{ rev min}^{-1}$  for 30 min. The resulting mixture was then centrifuged and analysed by the colorimetric Griess reaction method of Miranda *et al.* [16]. This is referred to as the standard lab method with KCl/ $\text{H}_2\text{O}$  extraction.

### 2.6. $\text{NO}_3^-$ Determination by Dual Wavelength UV Spectroscopy

$\text{NO}_3^-$  in the standard KCl and  $\text{H}_2\text{O}$  extracts were also analysed with dual wavelength UV spectroscopy at 205 nm and 300 nm using the method described in Edwards *et al.* [19].  $\text{NO}_3^-$  absorbs strongly at 205 nm, however, dissolved organic matter (DOM) also absorbs strongly at this wavelength. To compensate for this, the DOM can also be measured at 300 nm, where no  $\text{NO}_3^-$  absorption occurs, and the relationship between the DOM absorbance at 205 and 300 nm can be incorporated into a traditional  $\text{NO}_3^-$  calibration curve to account for the DOM present as follows:

$$\text{DOM}_{205} = (2.841 \times \text{DOM}_{300}) - 0.0126 \quad (1)$$

where  $\text{DOM}_{205}$  = organic matter absorbance at 205 nm,  $\text{DOM}_{300}$  = organic matter absorbance at 300 nm. The  $\text{DOM}_{205}$  absorbance value is simply subtracted from the sample reading prior to calculating  $\text{NO}_3^-$  from the standard curve. This method was originally developed for testing natural waters and to our knowledge has not been used for  $\text{NO}_3^-$  determination of soil extracts.

### 2.7. Evaluation of the Methods across a Broad Range of Soils

A diverse range of different soils ( $n = 23$ ) were sampled from within a  $10 \text{ km}^2$  radius of the Henfaes Research Station. The samples were analysed using the ISE rapid test method ( $n = 23$ ), the standard lab method with  $\text{H}_2\text{O}$  extraction ( $n = 23$ ) and UV spectroscopy ( $n = 16$ ) as described above. These soils were not spiked with  $\text{NO}_3^-$ . Background analysis of these soils can be found in Table 2.

**Table 2.** Background soil analysis for the broad range of soil sampled from within a  $10 \text{ km}^2$  radius of the Henfaes Research Station. EC = Eutric Cambisol, DG = Dystric Gleysol, HP = Haplic Podzol.

| Sample | Soil | pH   | EC ( $\mu\text{S cm}^{-1}$ ) | Moisture content ( $\text{g g}^{-1}$ ) | Organic matter (%) |
|--------|------|------|------------------------------|--|--------------------|
| 1      | DG   | 6.04 | 19.3                         | 0.33                                   | 5                  |
| 2      | EC   | 6.04 | 13.9                         | 0.38                                   | 7                  |
| 3      | HP   | 5.85 | 6.4                          | 0.51                                   | 10                 |
| 4      | HP   | 6.07 | 35.1                         | 0.32                                   | 5                  |
| 5      | HP   | 4.85 | 43.2                         | 0.52                                   | 18                 |
| 6      | HP   | 5.65 | 16.5                         | 0.81                                   | 17                 |
| 7      | HP   | 5.89 | 84.6                         | 0.36                                   | 9                  |
| 8      | EC   | 6.38 | 20.3                         | 0.19                                   | 7                  |
| 9      | EC   | 6.19 | 13.8                         | 0.17                                   | 5                  |
| 10     | EC   | 6.65 | 55.0                         | 0.17                                   | 8                  |

Table 2. Cont.

| Sample | Soil | pH   | EC ( $\mu\text{S cm}^{-1}$ ) | Moisture content ( $\text{g g}^{-1}$ ) | Organic matter (%) |
|--------|------|------|------------------------------|--|--------------------|
| 11     | EC   | 7.14 | 34.3                         | 0.18                                   | 6                  |
| 12     | EC   | 6.48 | 43.5                         | 0.29                                   | 7                  |
| 13     | EC   | 6.51 | 47.8                         | 0.39                                   | 12                 |
| 14     | EC   | 6.61 | 18.1                         | 0.25                                   | 4                  |
| 15     | EC   | 6.29 | 49.8                         | 0.52                                   | 10                 |
| 16     | EC   | 6.23 | 45.7                         | 0.30                                   | 8                  |
| 17     | DG   | 5.37 | 69.7                         | 0.63                                   | 8                  |
| 18     | HP   | 4.53 | 32.6                         | 0.37                                   | 5                  |
| 19     | EC   | 5.18 | 6.8                          | 0.30                                   | 11                 |
| 20     | EC   | 6.28 | 45.7                         | 0.24                                   | 6                  |
| 21     | EC   | 6.84 | 50.0                         | 0.20                                   | 6                  |
| 22     | EC   | 5.39 | 38.6                         | 0.39                                   | 5                  |
| 23     | EC   | 5.52 | 90.2                         | 0.46                                   | 10                 |

## 2.8. Statistical Analysis

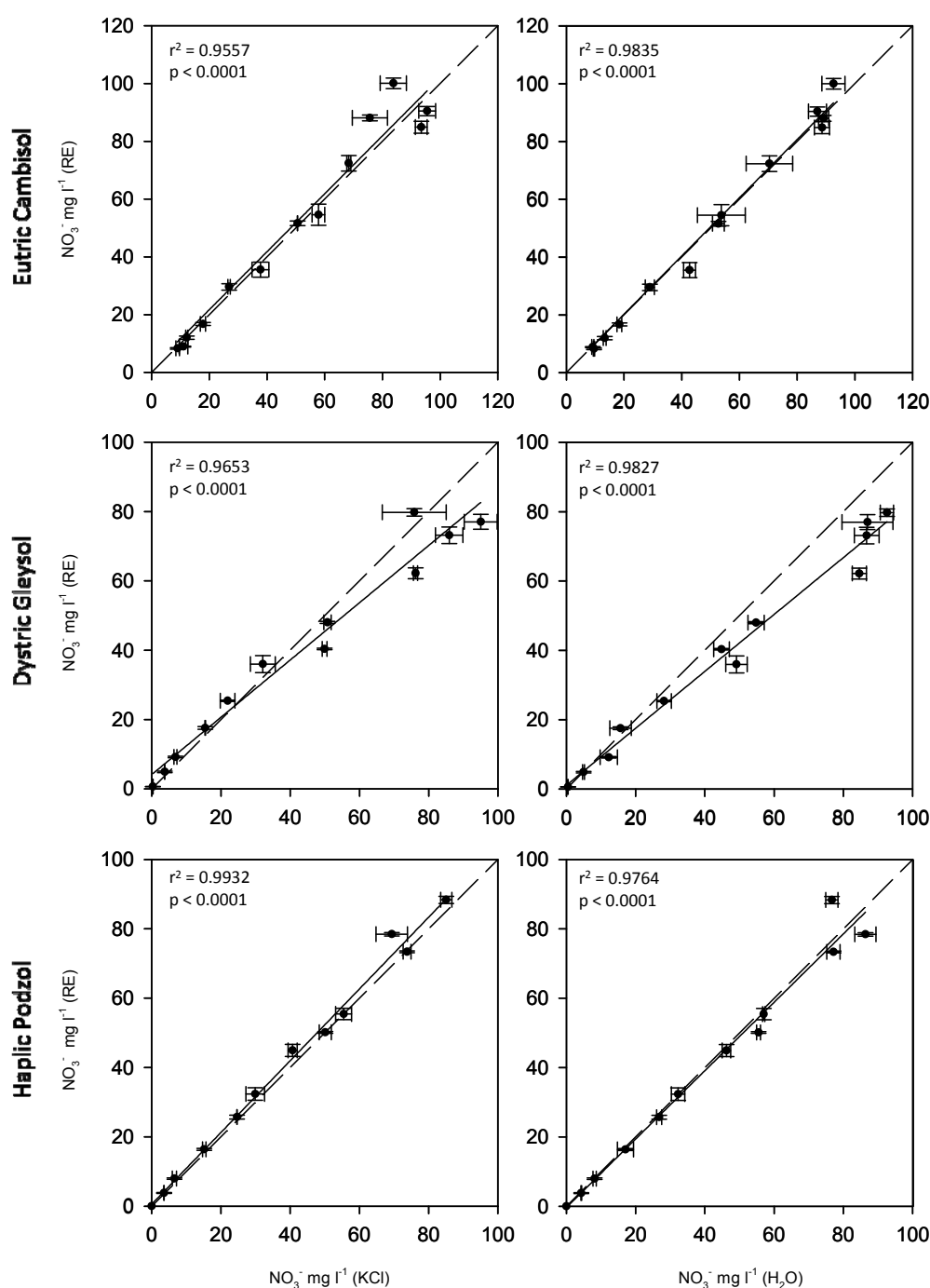
All concentrations given in this paper are reported in  $\text{mg NO}_3^- \text{ L}^{-1}$ . Linear regression analysis was undertaken using SigmaPlot v12.3 (Systat Software Inc., Hounslow, UK) and paired *t*-tests were undertaken with SPSS v20 (IBM Ltd., Portsmouth, UK).  $P < 0.05$  was used as the cut-off for statistical significance.

## 3. Results and Discussion

### 3.1. Efficiency of the Rapid Extraction Method

The efficiency of soil nitrate extraction from the three soil types using the rapid extraction method was tested by comparing it to standard soil extractions with 1 M KCl or  $\text{H}_2\text{O}$ . The standard lab method was used for the subsequent  $\text{NO}_3^-$  determination of all the extracts. Figure 1 shows an excellent correlation and a near 1:1 relationship between the rapid extraction method and the KCl or  $\text{H}_2\text{O}$  standard extraction method for the Eutric Cambisol and the Haplic Podzol, with no significant differences observed. In addition, there were no significant differences observed in  $\text{NO}_3^-$  extraction between the  $\text{H}_2\text{O}$  standard extraction and the 1 M KCl standard extraction for all three soil types, which shows that  $\text{NO}_3^-$  extraction using  $\text{H}_2\text{O}$  is acceptable. However, the efficiency of the rapid extraction method on the Dystric Gleysol was lower and was shown to be significantly different from the KCl standard extraction. The structure of the Dystric Gleysol was very poor so it is likely that shaking by hand for 2 min was not enough to allow complete dispersal of the soil particles. A similar problem may occur on heavy clay soils [15]. This may be rectified by increasing the shaking time of the extraction.

**Figure 1.** Comparison of the rapid extraction procedure (RE) with the standard 1 M KCl and H<sub>2</sub>O extracts. All extracts analysed for NO<sub>3</sub><sup>−</sup> using the standard lab method. The dotted line represents the theoretical 1:1 line for the two methods whilst the solid line represents the linear regression line describing the actual relationship between the two methods. Values represent means ± SEM (*n* = 3). The *r*<sup>2</sup> and *p* value from the regression analysis are shown for each graph.



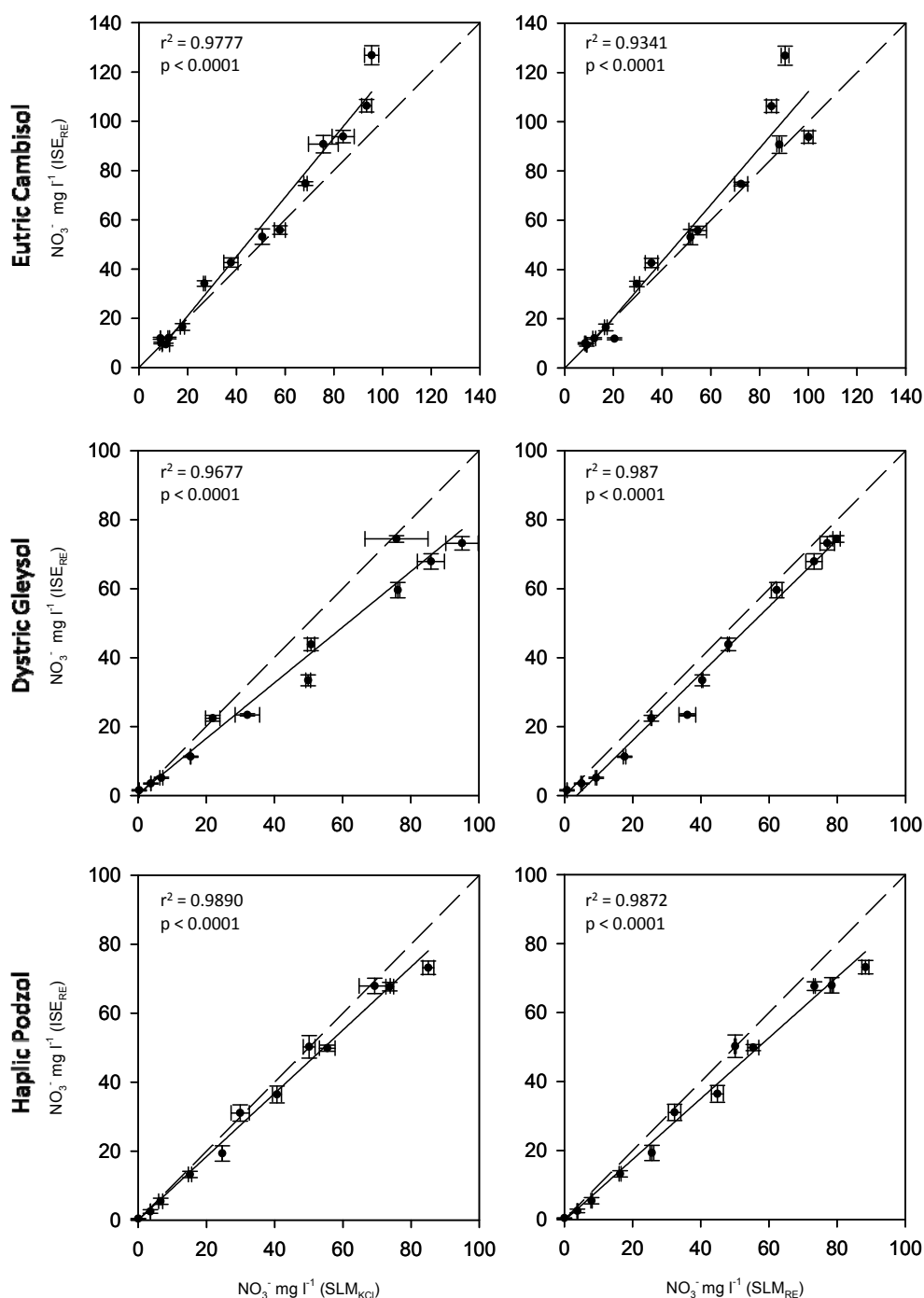


### 3.2. Comparison of the ISE Rapid Test with the Standard Lab Method

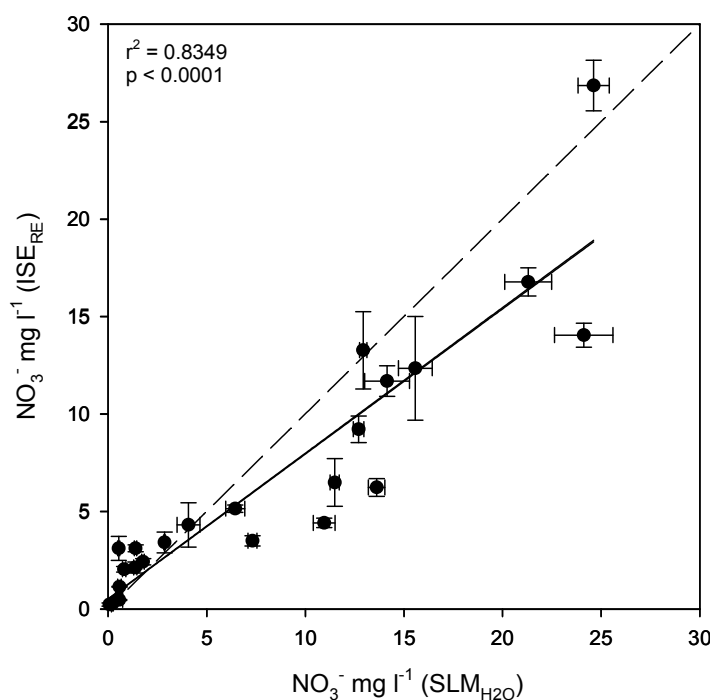
The ISE rapid test method was compared to the standard lab method with KCl extraction. Figure 2 shows a good correlation between the ISE rapid test and the standard lab method with KCl extraction for the determination of soil  $\text{NO}_3^-$  in all three soil types. However, significant differences were found between the two methods when applied to the Eutric Cambisol and Dystric Gleysol. Analysis of the rapid extraction extracts with the standard lab method for these soils showed no significant differences when compared to the ISE rapid test method. This suggests that the significant difference between the ISE rapid test method and the standard lab method with KCl extraction was due to either differences in extraction efficiency or natural soil variation, but not the performance of the ISE. We have already shown above that the efficiency of the rapid extraction method on the Dystric Gleysol is lower than for the other soil types, which would explain the reduced accuracy of the ISE rapid test for this soil.

Figure 3 shows the results of the ISE rapid test compared to the standard lab method with  $\text{H}_2\text{O}$  extraction for the determination of  $\text{NO}_3^-$  on a range of soils, which were not spiked with  $\text{NO}_3^-$ . At these lower  $\text{NO}_3^-$  concentrations, the ISE rapid test tends to underestimate the  $\text{NO}_3^-$  concentration, the correlation between the two methods was not quite as strong and there was a significant difference between them. In particular, below  $10 \text{ mg L}^{-1}$  there is a poor response of the ISE to changing concentration. The efficiency of the rapid extraction method was not ascertained for the soils used here and the low values of  $\text{NO}_3^-$  would exacerbate any reduction in extraction efficiency, although this is only likely to be an issue for the two Dystric Gleysols. In addition, the response of the ISE below  $10 \text{ mg L}^{-1}$  was non-linear and so as  $\text{NO}_3^-$  concentration decreased below this the resolution of the ISE was reduced. The accuracy in the non-linear phase may also be reduced by the calibration method. The software used calculates the calibration from the standards using a semi-logarithmic interpolation method. Essentially, this works by joining the calibration points with a straight line, which has obvious implications for a non-linear curve. Accuracy in the non-linear phase may therefore be increased by using more calibration points. Alternatively, a curve could be fitted to the calibration data using a simplified version of the Nicolsky–Eisenman equation [20].

**Figure 2.** Comparison of the ion selective electrode (ISE) rapid test ( $\text{ISE}_{\text{RE}}$ ) with the standard lab method—extractions in KCl ( $\text{SLM}_{\text{KCl}}$ ) and rapid extraction procedure ( $\text{SLM}_{\text{RE}}$ )—for  $\text{NO}_3^-$  determination in three soils amended with increasing amounts of  $\text{NO}_3^-$ . The dotted line represents the theoretical 1:1 line for the two methods whilst the solid line represents the linear regression line describing the actual relationship between the two methods. Values represent means  $\pm$  SEM ( $n = 3$ ). The  $r^2$  and  $p$  value from the regression analysis are shown for each graph.



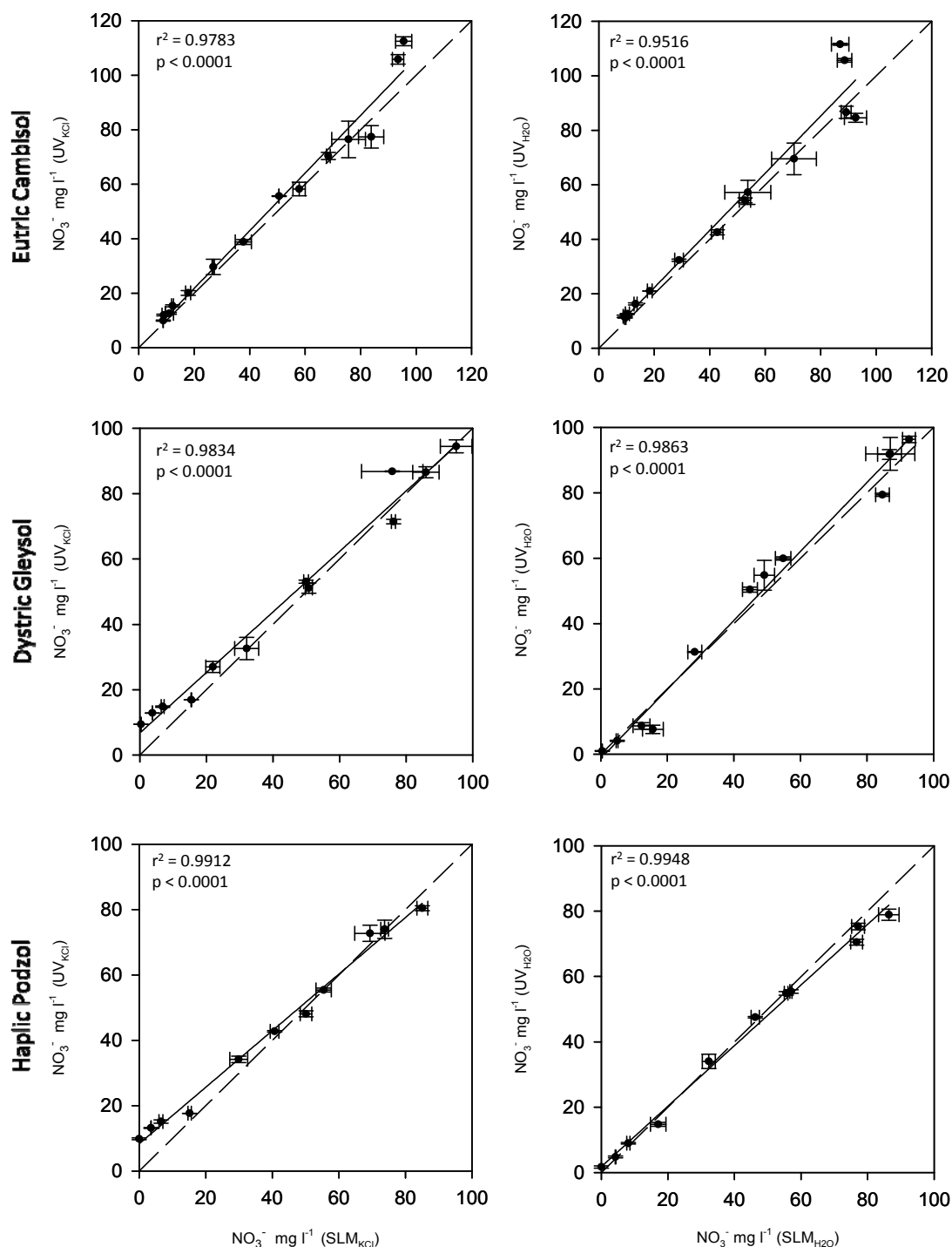
**Figure 3.** Comparison of the ISE rapid test with the standard lab method—extractions in H<sub>2</sub>O (SLM<sub>H2O</sub>)—for NO<sub>3</sub><sup>−</sup> determination across a broad range of agricultural soils ( $n = 23$ ). The dotted line represents the theoretical 1:1 line for the two methods whilst the solid line represents the linear regression line describing the actual relationship between the two methods. Values represent means  $\pm$  SEM ( $n = 3$ ). The  $r^2$  and  $p$  value from the regression analysis are shown.



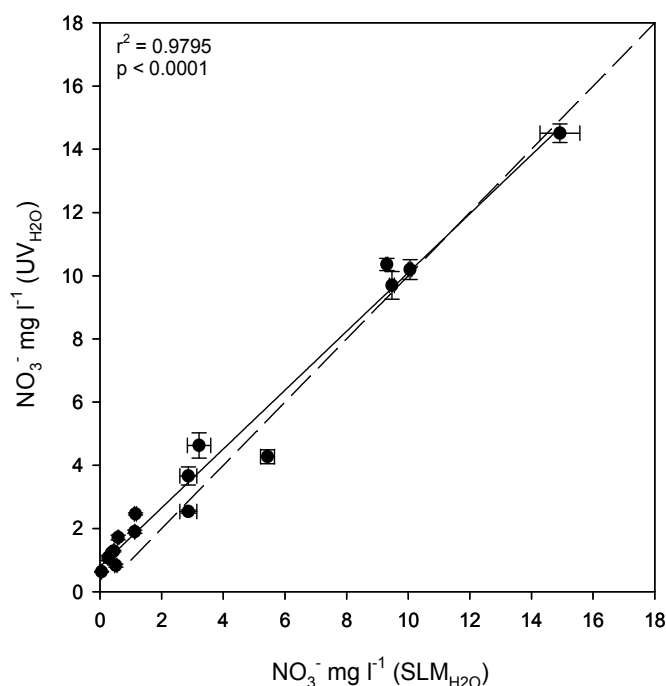
### 3.3. Comparison of UV Spectroscopy with the Standard Lab Method

Figures 4 and 5 show an excellent correlation and a near 1:1 response between the standard lab method and UV spectroscopy method for the determination of NO<sub>3</sub><sup>−</sup> in H<sub>2</sub>O and KCl soil extracts. The response of UV spectroscopy to pure solutions of NO<sub>3</sub><sup>−</sup> was linear from 0.05–12.5 mg L<sup>−1</sup> (compared to 0.2–50 mg L<sup>−1</sup> with the standard lab method). Consequently, most of the extracts required a 1:10 (v/v) dilution prior to NO<sub>3</sub><sup>−</sup> determination. No significant differences were found between the methods using H<sub>2</sub>O extraction for all three soil types and for KCl extraction with the Haplic Podzol. However, significant differences were found between the standard lab method and the UV spectroscopy for KCl extractions from the Eutric Cambisol and Dystric Gleysol. A closer look at Figure 4 shows that it is only the three lowest concentrations that appear to deviate significantly from the 1:1 regression line. These were the only samples, extracted in 1 M KCl which were not diluted prior to UV analysis, which suggests that the error is due to interference from the 1 M KCl. Edwards *et al.* [19], found no interference from saline constituents although they did not use solutions as strong as 1 M. Figure 5 shows that unlike the ISE rapid test there was no loss of accuracy at low concentrations. However, here only the analytical methods are being compared and both methods are using the same extracts, whereas with the ISE rapid test different extractions are used leading to variation in both the extraction efficiency and natural soil variation.

**Figure 4.** Comparison of UV spectroscopy—extractions in KCl ( $UV_{KCl}$ ) and  $H_2O$  ( $UV_{H_2O}$ )—with the standard lab method—extractions in KCl ( $SLM_{KCl}$ ) and  $H_2O$  ( $SLM_{H_2O}$ )—for  $NO_3^-$  determination in three soils amended with increasing amounts of  $NO_3^-$ . The dotted line represents the theoretical 1:1 line for the two methods whilst the solid line represents the linear regression line describing the actual relationship between the two methods. Values represent means  $\pm$  SEM ( $n = 3$ ). The  $r^2$  and  $p$  value from the regression analysis is shown for each graph.



**Figure 5.** Comparison of UV spectroscopy—extractions in H<sub>2</sub>O (UV<sub>H2O</sub>)—with the standard lab method—extractions in H<sub>2</sub>O (SLM<sub>H2O</sub>)—for NO<sub>3</sub><sup>−</sup> determination across a broad range of agricultural soils ( $n = 16$ ). The dotted line represents the theoretical 1:1 line for the two methods whilst the solid line represents the linear regression line describing the actual relationship between the two methods. Values represent means  $\pm$  SEM ( $n = 3$ ). The  $r^2$  and  $p$  value from the regression analysis is shown.



### 3.4. Evaluation of ISEs for Soil Nitrate Determination

The ISE rapid test procedure was conducted on three contrasting soil types (Eutric Cambisol, Dystric Gleysol and Haplic Podzol), which together represent the major agricultural soil types in the UK. In comparison to the internationally recognised standard laboratory method, the results clearly showed that current ISE technology combined with two min manual H<sub>2</sub>O soil extractions has the potential to be used by farmers as an on-farm rapid-diagnostic test. However, the accuracy of the rapid test procedure decreased when testing a Dystric Gleysol and at low NO<sub>3</sub><sup>−</sup> concentrations (*i.e.*, below 10 mg L<sup>−1</sup>). There was also a problem with the electrode durability. During this experiment, three sets of ISEs were used. The second set was discarded when it began to show a large erratic response and a subsequent failure to stabilise. This type of malfunction is likely to be due to a failure of the electronics and can easily be spotted. The first and third set suffered from a loss of sensitivity at the lowest concentrations and the third set showed physical degradation manifest by a bulging membrane. The electrodes were discarded when they could not be calibrated correctly at the concentrations that were being determined. However, a subtle loss of sensitivity at lower concentrations or changes in the calibration parameters may not be spotted by a layman and could cause significant error if the calibration was not adjusted. A more rugged sensor housing design may improve this lack of durability alongside changes in the sensor chemistry

(e.g., inclusion of protective membranes). In addition, electrodes can also be constructed incorporating the latest improvements in ion sensor membrane design [21].

### 3.5. Evaluation of UV Spectroscopy for Soil Nitrate Determination

The results show clearly that dual wavelength UV spectroscopy can be used to accurately determine the concentration of  $\text{NO}_3^-$  in extracted soil solution. When combined with the rapid extraction method it has the potential to be used as an on-farm rapid test providing a hand-held UV spectrometer is available and the extracts are filtered or centrifuged. Dual wavelength UV spectroscopy is able to determine the  $\text{NO}_3^-$  concentration between 0.35 and 17.7  $\text{mg L}^{-1}$  [19]. This means that extractions from soils with high  $\text{NO}_3^-$  input may need to be diluted before measurement. This was the case in our study for the majority of the samples. The results suggest that the 1 M KCl extractant causes some interference to the measurement. Diluting KCl extractions 1:10 (v/v) appears to prevent the interference that occurs due to 1 M KCl. Extracting with  $\text{H}_2\text{O}$  also solves the problem of the 1 M KCl interference and the results show there is no difference in  $\text{NO}_3^-$  extraction using 1 M KCl or  $\text{H}_2\text{O}$ . In addition, distilled water can be readily purchased by most farmers in comparison to KCl solutions. The advantage of using an ISE over UV spectroscopy for an on-farm rapid test is that ISE's can be used in soil slurry so no filtration or centrifugation needs to be carried out. However, UV spectroscopy offers better resolution at very low concentrations due to the linear nature of its response and is likely to be more durable. In addition, field portable UV spectrometers are now readily available. This approach, however, is readily suited to the evaluation of nitrate in agricultural drainage waters.

### 3.6. Implications for Fertiliser Application Guidelines

Extraction of  $\text{NO}_3^-$  from the soil and its subsequent determination does not provide all the information required to produce an agronomic relevant result. For the results to be meaningful for agricultural extension purposes, they must be up-scaled to units of  $\text{kg ha}^{-1}$ . This requires determination of the bulk density and moisture content of the samples. Schmidhalter [15] developed a method of determining both parameters using a simple in-field method requiring only standard bulk density cylinders, a graduated measuring cylinder and a solar powered balance.

For farmers to implement rapid soil  $\text{NO}_3^-$  testing, they must be convinced of the benefit as the process requires both time and money. When soil is sampled and sent for laboratory testing it can be analysed for a range of macro- and micro- nutrients. In the UK, fertiliser additions, as prescribed by RB209, require a calculation or estimation of SMN, which includes both  $\text{NH}_4^+$  and  $\text{NO}_3^-$ . This rapid test would only determine soil  $\text{NO}_3^-$  concentration, and although nitrogen is fundamental to plant growth, it is not always the limiting nutrient.  $\text{NO}_3^-$  differs from other nutrients in that its concentration varies greatly both spatially and temporally, which is the main reason that farmer-operated  $\text{NO}_3^-$  rapid tests performed through the growing season may improve fertiliser management. This spatial and temporal variation does however pose a challenge as to determining the optimum sampling regime. Further, farmers need to have relevant and simple decision support systems so that collected data are interpreted correctly and can be implemented into a meaningful fertiliser strategy. Along with improving the technology, further work is therefore needed so that rapid soil  $\text{NO}_3^-$  can be adopted by industry as a way to optimise nutrient use efficiency.

#### 4. Conclusions

This work shows that ISEs can offer a reasonably accurate and rapid way of determining  $\text{NO}_3^-$  concentration in soil slurries. This can be combined with a rapid extraction procedure using  $\text{H}_2\text{O}$  where the soil is shaken by hand for 2 min. For poorly structured or heavier clay soils a longer shaking time may be required. There is the potential for ISEs to be used by farmers for an on-farm rapid test; however, practicality issues and methods for integrating the data into a management plan may reduce its uptake. UV spectroscopy offers a similarly rapid and reagentless method of  $\text{NO}_3^-$  determination. Compared to ISE, it offers a lower detection limit and a greater resolution at low concentration—below  $10 \text{ mg L}^{-1}$ —but samples with a concentration greater than  $18 \text{ mg L}^{-1}$  will need to be diluted for accurate determination. The technology is likely to be more durable and less prone to error than ISEs. However, the cost of the technology is likely to be greater and samples will require filtering or centrifuging prior to measurement. For rapid tests to be used by the industry, farmers must be convinced of the cost-benefits and have a suitable decision support mechanism in place to turn the measurements into a fertiliser application plan.

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